# **High-entropy oxide (CeGdHfPrZr)O2 nanoparticles as reusable**   $\mathbf{1}$

# **photocatalyst for wastewater remediation**  $\overline{2}$

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# $\overline{7}$ **Highlights:**

- 8 Single-Phase (CeGdHfPrZr) $O_2$  was synthesized using a hydrothermal technique.
- $(CeGdHfPrZr)O<sub>2</sub>$  showed better photocatalytic activity at alkaline pH values. 9
- · Hydroxyl radical is responsible for MB degradation.  $10\,$
- Better photocatalytic recyclability is observed without any phase transformation. 11
- e for MB degradation.<br>lity is observed without are • (CeGdHfPrZr)O<sub>2</sub> appears to be an attractive photocatalytic material for wastewater 12 remediation. 13

#### **Abstract:**  $14\,$

15 High-entropy materials (HEM) play a significant role in current scientific research and are 16 characterized by their complexity, which makes them the next generation of nanomaterials. The present study investigates the synthesis of  $(CeGdHfPrZr)O<sub>2</sub>$  high-entropy oxide 17 nanoparticles using a hydrothermal technique. Various characterization techniques, such as X-18 19 ray diffraction (XRD), and scanning electron microscopy (SEM) were used to investigate the 20 structural properties, while UV-visible spectroscopy was used to investigate the optical properties. The results indicate the formation of a single-phase cubic fluorite-high-entropy 21

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oxide system with a mean crystallite size of 5.4 nm. The optical bandgap of  $(CeGdHfPrZr)O<sub>2</sub>$ 22 nanoparticles is found to be 2.01 eV. The photocatalytic activity of the synthesized oxide 23 nanoparticles was assessed using methylene blue (MB) dye as a model pollutant. In the current 24 study, we examined and discussed the effect of various photocatalytic parameters on the 25 degradation of MB dye. The results showed that the  $(CeGdHfPrZr)O<sub>2</sub>$  nanoparticles had high 26 27 photocatalytic activity and were found to be dependent on various parameters. A higher 28 photocatalyst concentration impedes the degradation kinetics and, as a result, limits photon penetration into the reaction solution. Similarly, increased hydroxyl radical generation at a 29 basic pH improved MB dye degradation. In addition, based on the band positions and radical 30 scavenging study, hydroxyl radicals are responsible for the degradation of dye. 31  $(CeGdHfPrZr)O<sub>2</sub>$  nanoparticles can be used as a promising catalyst for the photocatalytic 32 Photocatalyst; Methylen degradation of organic pollutants. 33

## **Keywords:** 34

High-entropy oxide; Solid solution; Photocatalyst; Methylene Blue; Structural stability 35

#### 36 **1. Introduction:**

37 Research on a new class of nanomaterials referred to as "High-entropy materials" (HEM) has received considerable attention in recent years. A HEM system is unlikely to differ from a 38 conventional doped system in many aspects. Typically, HEMs are formed by combining five 39 40 or more principal elements to achieve complete solubility and, as a result, form a solid solution. 41 Therefore, HEMs are extremely stable and versatile, which makes them suitable for a wide range of applications. This in turn contributes to the creation of a new paradigm in scientific 42 research regarding HEMs to create materials with unique physical and chemical properties [1- 43 3]. High-entropy oxide (HEO) is one of the most intriguing systems that have been shown to 44 exhibit better thermal properties [4, 5], optical [6], photoluminescence [7], dielectric [8, 9], 45

thermoelectric [10], magnetic [11-13], etc. TiZrNbHfTaO<sub>11</sub> oxide, for example, outperformed 46 typical photocatalysts such as anatase  $TiO<sub>2</sub>$  and  $BiVO<sub>4</sub>$ , and was comparable to P25  $TiO<sub>2</sub>$  in 47 terms of photocatalytic CO<sub>2</sub> conversion [14]. Furthermore,  $Pr_{1/6}La_{1/6}Da_{1/6}Sr_{1/6}Ca_{1/6}CoO_{3-\delta}$ 48 is reported as a high-performance bifunctional air electrode for reversible proton ceramic 49 electrochemical cell compared to advanced air electrodes reported so far[15]. These findings 50 51 demonstrate the advantages of designing high-entropy materials in comparison to pure or 52 doped oxide system.

53 Moreover, their tuneable structural and chemical properties make them ideal for use in a wide range of practical applications, including energy storage [8, 16], sensors [17-19], thermal 54 barrier coatings [20], catalysis [21], electrocatalysts [22], and photocatalysts [6, 23]. Despite 55 the attention given to other functional applications, few research has been conducted on using onal applications, few reservises.<br>
is, thus leaving the area leading the original catalyst for studying the original catalyst for studying the original catalogue. 56 HEO in the field of photocatalysis, thus leaving the area largely unexplored. Edalati *et al.*  57 produced a TiZrNbTaWO<sub>12</sub> photocatalyst for studying the oxygen evolution reaction [24]. Yu 58 *et al.* studied the photocatalytic reduction of Cr(VI) using  $A_{32}Ti_8Sn_8Nb_4Ta_4Me_8O_{96}$  (A=Ba, Sr; 59 Me=Fe, Ga) perovskite structure and found that the system containing Fe has the best 60 61 photocatalytic activity [25]. Anandkumar *et al.* studied the photocatalytic conversion of Cr(VI) and degradation of methylene blue using Gd<sub>0.2</sub>La<sub>0.2</sub>Ce<sub>0.2</sub>Hf<sub>0.2</sub>Zr<sub>0.2</sub>O<sub>2</sub> and 62  $Gd_{0.2}La_{0.2}Y_{0.2}Hf_{0.2}Zr_{0.2}O_2$  high-entropy oxide. Both the oxide systems display good Cr(VI) 63 64 reduction and better methylene blue degradation [6].

65 Considering the significant gap left in the field of high-entropy photocatalysts, wastewater remediation is of particular interest to us. A competitive economic race has led to rapid 66 industrialization, which in turn has resulted in an increase in wastewater discharge [26-29]. 67 Despite the existence of strict regulations, the possibility of industrial effluents contaminating 68 69 fresh water remains uncontrollable, and serious action must be pursued. An increase in 70 monitoring and enforcement efforts, as well as improvements in wastewater treatment 71 processes, may be necessary to reduce water pollution. Various contaminants such as methyl orange, methylene blue, and other heavy metals such as Cr, Cd, Hg, etc. have been discharged 72 from different industries like leather, textiles, mining, cosmetics, etc[6]. The textile and paper 73 industries use more than 10,000 synthetic dyes and pigments, causing significant negative 74 environmental and health impacts[27]. Water bodies are at risk from the direct discharge of 75 76 industrial effluents which pose serious ecotoxicological threats. Several health issues for 77 human beings as well as for wildlife as a result of ingesting contaminated water have been identified [6]. In order to protect humans and wildlife, it is imperative to reduce the amount of 78 these pollutants in our environment. 79

To neutralize the harmful chemicals, present in wastewater discharge, the advanced 80 81 oxidation process (AOP) through heterogeneous photocatalysis is one of the most viable and heterogeneous photocataly<br>
nating industrial dyes usin<br>
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ates an electron-hole pair<br>
th levels of ROS, which are 82 eco-friendly techniques for eliminating industrial dyes using reactive oxygen species (ROS) [30, 31]. The photocatalytic action is usually carried out using either natural sunlight or a 83 UV/visible light source that generates an electron-hole pair on the photocatalyst surface [32]. 84 This leads to the generation of high levels of ROS, which are responsible for neutralizing dye 85 86 molecules. The AOP process is cost-effective, and efficient and the degradation product resulting from this process is least harmful to the environment and soil. Sudarsan *et al.* 87 synthesized a copper ferrite nanocomposite by solvothermal technique from discarded printed 88 89 circuit boards and used it as a photocatalyst for the degradation of Congo red dye. The prepared copper ferrite nanocomposite could effectively decompose 96.19% of dye within 120 min [33]. 90 Doped systems such as Sr doped BaTiO<sub>3</sub> show better performance for the degradation of 91 rhodamine B with a degradation rate of 92.66 % within 18 min[34]. 92

Similarly, several heterojunctions have been explored in order to degrade methylene 93 blue and toxic metals such as Cr(VI) [35-38]. A number of nanomaterials have been reported 94 95 to degrade different dyes [34, 39]. Ceria-based nanocomposite  $(Ce_{0.8}Y_{0.2}O_{2.5}-Ce_{0.8}Sm_{0.2}O_{2.5})$ 

nanoparticles fabricated by the chemical co-precipitation technique show a degradation 96 efficiency of 93%. Similarly, NiO-Ce<sub>0.9</sub>Y<sub>0.1</sub>O<sub>2-δ</sub>–Ce<sub>0.9</sub>Sm<sub>0.1</sub>O<sub>2-δ</sub> nanoparticle photocatalyst was 97 used to degrade congo red dye[40]. As a result, we investigate a new photocatalytic material 98 for the degradation of methylene blue dye to ensure the environmental effects can be 99 minimized. 100

To fill this gap, our scientific work seeks to develop new high-entropy photocatalysts 101 that can effectively remove contaminants from wastewater. In this present study, we have 102 synthesized (CeGdHfPrZr) $O_2$  nanoparticles using a hydrothermal route using NaOH and 103 Polyvinylpyrrolidone. Further, we have investigated the photocatalytic activity of 104  $(CeGdHfPrZr)O<sub>2</sub>$  nanoparticles for the photocatalytic degradation of methylene blue dye. An 105 106 in-depth investigation of the photocatalytic degradation of MB dye has been carried out by ocatalytic degradation of<br>prameters and is discussed<br>plications for HEOs in the varying different photocatalytic parameters and is discussed in detail. As a result of this study, 107 we may be able to reveal novel applications for HEOs in the area of photocatalysis. 108

**2. Experimental section** 109

## 110 **2.1. Materials and reagents**

Cerium (III) nitrate hexahydrate  $(Ce(NO<sub>3</sub>)<sub>3</sub>•6H<sub>2</sub>O)$ , zirconium (IV) nitrate dihydrate 111 (Zr(NO3)2•2H2O), hafnium (IV) oxychloride (HfOCl2•8H2O), gadolinium (III) nitrate 112 hexahydrate  $(Gd(NO<sub>3</sub>)<sub>3</sub>•6H<sub>2</sub>O)$ , praseodymium (III) nitrate hexahydrate  $(Pr(NO<sub>3</sub>)<sub>3</sub>•6H<sub>2</sub>O)$ , 113 methylene blue, hydrogen peroxide (H2O2), hydrochloric acid (HCl), sodium hydroxide 114 115 (NaOH), Polyvinylpyrrolidone (PVP) were purchased from China. All the materials were marked with the analytical grade and were used as received without any purification. The 116 deionized (DI) water was used for the synthesis. 117

# **2.2. Synthesis of (CeGdHfPrZr)O2 oxide nanoparticles** 118

A simple hydrothermal synthesis was employed for the synthesis of high-entropy oxide 119 nanoparticles. Initially, respective metal salts (0.001 M of each metal salt) were weighed and 120 dissolved in 80 ml of DI water, followed by stirring at room temperature until all the salts were 121 dissolved completely. Then PVP was added to the above transparent solution and stirred until 122 the PVP dissolved completely. The molar ratio of PVP to metal salts is 1:2. Then, the required 123 124 amount of NaOH solution was added dropwise until a pH of 10 was achieved, and stirring was 125 continued for another 10 min. The reaction mixture was then transferred to a 100 ml hydrothermal reactor vessel and kept in a preheated oven at 80 °C for 6 h. After the completion 126 of the reaction, the reactor vessel was cooled down to room temperature and washed with DI 127 water multiple times until the pH of the solution was 7. The obtained precipitate was dried in 128 an oven at 110 °C for 12 h to evaporate the solvent. The dried precipitate was ground in an 129 air calcination in a muffle agate mortar pestle, followed by air calcination in a muffle furnace at 500  $^{\circ}$ C for 4 h. The 130 obtained **(**CeGdHfPrZr)O2 powder was used for further characterization and photocatalytic 131 investigation. 132

#### **2.3. Characterization** 133

The crystallinity of the prepared HEO was studied using XRD (Rigaku Ultima IV, 134 Rigaku) using Cu-K<sub>α</sub> radiation ( $\lambda = 1.54$  Å) in the range of 20-80° with a scan speed of 1° per 135 min. Rietveld refinement was done using the Fullprof program, and Williamson-Hall (W-H) 136 analysis was done to assess the contribution of crystallite size and the lattice strain to XRD 137 peak broadening. FESEM images were captured using a JEOL (JEOL JSM-7001F, JEOL) 138 microscope operated at a potential of 20 kV. Similarly, the elemental quantification was done 139 by energy-dispersive x-ray spectroscopy (EDS) equipped with a field emission scanning 140 electron microscope (JEOL JSM-7001F, JEOL). The HRTEM image and SAED pattern were 141 captured using a JEOL microscope (JEOL JEM-2100, JEOL) operated at an acceleration 142 voltage of 175 kV. The optical properties were investigated using a UV-Visible 143

spectrophotometer (Shimadzu UV-2700) in both absorbance and reflectance modes. Fourier-144 transform infrared spectroscopy (FTIR) studies were performed on the powders using the 145 Shimadzu IRAffinity-1S. Sample preparation for the FTIR studies involves the preparation of 146 the pellet by mixing the powder photocatalyst and KBr (1:100 weight ratio). Data was collected 147 in transmittance mode from 400-4000 cm<sup>-1</sup>. 148

## **2.4. Photocatalytic degradation of Methylene Blue dye** 149

The degradation of MB dye was performed in the presence of  $(CeGdHfPrZr)O<sub>2</sub>$  catalyst 150 under UV light as an irradiation source. Initially, 50 mg of HEO catalyst was added to 100 ml 151 of 20 mgL<sup>-1</sup> MB solution, followed by the addition of 200 µl of  $H_2O_2$  solution to boost the 152 generation of hydroxyl radicals in the system. To initiate the adsorption-desorption 153 was magnetically stirred for<br>
t and MB dye is irradiated<br>
. Meanwhile, 3 mL of the<br>
trifuged at a speed of 120 154 equilibrium, the reaction solution was magnetically stirred for 60 min in the dark. After 60 min, the solution containing the catalyst and MB dye is irradiated with UV light ( $\lambda = 402$  nm) which 155 initiates the photocatalytic activity. Meanwhile, 3 mL of the reaction solution was taken out at 156 a specific interval of time and centrifuged at a speed of 12000 rpm for 5 min to separate the 157 catalyst from the reaction solution. The degradation of MB dye in the centrifuged solution was 158 monitored using a UV-visible spectrophotometer from 400 nm to 800 nm. The solution 159 160 containing MB dye was responsible for an absorption maximum of around 664 nm in the absorption spectra. Based on the intensity at 664 nm, the photocatalytic reaction kinetics were 161 calculated. Several reaction parameters have been investigated, such as photocatalyst 162 concentration (100, 250, 500, and 1000 mgL<sup>-1</sup>), MB concentration (10, 25, 50, and 100 mgL<sup>-</sup> 163 164  $1$ ), and pH (2.5, 4.4, 8.6, and 10.5). In addition, control experiments have been carried out individually in the absence of a photocatalyst and irradiation source. 165

#### 166 **3. Results and discussions**

The phase purity of calcined high-entropy **(CeGdHfPrZr)** $O<sub>2</sub>$  oxide nanoparticles were 167 examined by XRD. The X-ray diffraction pattern of the HEO oxide nanoparticle is shown in 168 Fig. 1. All the reflections are indexed to the (111), (200), (220), (311), (222), (400), and (331) 169 planes of the cubic fluorite structure. The obtained patterns are similar to those of pure  $CeO<sub>2</sub>$ 170 oxide with ICDD card number: 01-073-6318 [41]. No additional reflections are observed in the 171 calcined **(CeGdHfPrZr)**O<sub>2</sub> sample, which confirms the absence of other phases or impurities. 172 As a result, the synthesized HEO is a single-phase material and has a cubic fluorite structure. 173 The HEO exhibits a shift in the x-ray reflection at higher angles when compared with the 174 standard pattern, indicating the emergence of a unique lattice parameter as a result of mixing 175 176 five principal constituents.

tveld refinement was performed to the Rietveld ref To investigate further, Rietveld refinement was performed to fit the x-ray diffraction 177 178 pattern to calculate the lattice parameter. The Rietveld refinement was fitted using a cubic fluorite structure using a space group Fm-3m (225) and the corresponding fit is displayed in 179 Fig. 1b and the resulting fitting parameter is shown in Table S1. The Rietveld refinement fitting 180 parameters indicate that the synthesized HEO is single phase, and the calculated lattice 181 parameter is 5.331 Å. To estimate the crystallite size and lattice strain in the HEO oxide, the 182 Williamson-Hall (W–H) method was employed (Fig. S2). Accordingly, an average crystallite 183 size of 5.4 nm with a lattice strain of 1.78% is obtained. For HEO oxide systems, increased 184 lattice strain results from the incorporation of different-sized metal cations into a single crystal 185 structure. This increased lattice strain is one of the unique properties of a high-entropy oxide 186 system (severe lattice distortion effect) [1]. 187



**Fig. 1.** (a) X-ray diffraction pattern of synthesized HEO oxide nanoparticles and (b) Rietveld 189 190 refinement fit of HEO.

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To investigate the morphological features and elemental distribution of the synthesized 191 HEO oxide nanoparticles, FESEM was performed. Fig. 2(a) and (b) reveal agglomerated 192 From the hydrothermal syn<br>transferred to the distribution of the three candomly distributed<br>nettal cations and their correctal cations and their corrected spherical nanoparticles resulting from the hydrothermal synthesis route. In addition, to probe 193 194 the elemental distribution, elemental mapping was done and displayed in Fig. 2(c-i). It is observed that all five principal elements are randomly distributed without any segregation. This 195 confirms the complete mixing of metal cations and their corresponding EDS spectra, as shown 196 in Fig. 2(j). As can be seen in Table S2, the chemical composition estimated from the EDS 197 spectrum of the HEO confirms that all five elements are present in nearly equimolar amounts. 198

spectral absorbance. This eventually decreases the bandgap of our synthesized HEO oxide 225 nanoparticles. The indirect bandgap energies of HEO were calculated using the Kubelka-Munk 226 (K-M) function, F(R) by plotting hv versus (F(R)\* hv) (1/2) [7]. The band gap values were 227 calculated by extrapolating the linear part of the graph  $(F(R)*hv)$  (1/2) = 0. Using this equation, 228 a bandgap of 2.01 eV is obtained. The low bandgap value obtained for this **(**CeGdHfPrZr)O2 229 composition is however important for photocatalytic applications as it will improve the spectral 230 absorption in the visible region [46, 47] and the visual appearance of the HEO photocatalyst is 231 232 shown in the inset of Fig. 4.





**Fig. 4.** UV-visible absorption spectrum of  $(CeGdHfPrZr)O<sub>2</sub>$  oxide powder and its physical 234 appearance. Inset shows the bandgap calculation using the K-M function. 235

XPS was carried out to investigate the chemical environment of individual elements 236 and is shown in fig.S2. Based on the wide energy survey spectrum, it was found that Ce3d, 237 Gd4d, Hf4f, Zr3d, Pr3d, and O1s have characteristic peaks that corresponds to the existence of 238 Ce, Gd, Hf, Zr, Pr, and O elements, respectively. The high-resolution Ce 3d XPS spectrum 239 contains peaks at 885.28 eV, and 903.48 eV corresponding to  $Ce^{3+}$  while peaks at 882.28 eV, 240 888.28 eV, 898.19 eV, 900.58 eV, 907.38 eV, and 916.68 eV are attributed to  $Ce^{4+}$  species[48]. 241

As a result, the synthesized photocatalyst contains defects in terms of oxygen vacancies, which 242 is beneficial in photocatalytic investigations[6]. Similarly, two peaks whose binding energy 243 located at 147.18 eV (Gd<sup>3+</sup> 4d<sub>3/2</sub>) and 142.48 eV (Gd<sup>3+</sup> 4d<sub>5/2</sub>) corresponds to Gd<sup>3+</sup>. In case of 244 Hf4f spectra, binding energies at 17.08 eV, 18.58 eV correspond to the spin orbit doublets Hf 245  $4f_{7/2}$ , and Hf  $4f_{5/2}$ [49]. The Pr3d spectrum contain two peaks having binding energies 928.18 246 eV and 932.28 eV indicating the presence of  $Pr^{3+}$  and  $Pr^{4+}$  respectively[50]. The binding 247 energies at 181.58 eV and 183.78 eV correspond to doublet Zr  $3d_{5/2}$ –Zr  $3d_{3/2}$  which is due to 248 the presence of  $Zr^{4+}[51]$ . The oxygen spectra contains three peaks at 532.58 eV, 531.08 eV, 249 529.18 eV denoting hydroxyl groups, adsorbed oxygen, and lattice oxygen[52]. Therefore, 250 based on the XPS studies, it is evident that  $(CeGdHfPrZr)O<sub>2</sub>$  high-entropy oxide contains 251 oxygen vacancies enabling faster photocatalytic performances. 252

# **3.1 Photocatalytic degradation of methylene blue dye** 253

of methylene blue dye 254 To test the photocatalytic activity of synthesized HEO, a standard methylene blue dye is used as a model industrial dye, and UV light is used as an irradiation source. A well-known 255 pollutant, methylene blue, is used in a variety of industries for a variety of purposes. In order 256 to eradicate such hazardous pollutants from water bodies, it is vital to develop advanced 257 oxidation strategies by designing novel photocatalysts. Here, HEO is investigated as a novel 258 photocatalyst for MB degradation. 259

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# **3.1.1 Effect of catalyst loading**

To investigate the effect of catalyst loading on the photocatalytic property, different 261 concentrations of HEO catalyst, such as 100 mg/L, 250 mg/L, 500 mg/L, and 1000 mg/L were 262 added to the 20 mg/L MB solution, and its photocatalytic studies were monitored using a UV-263 visible spectrophotometer, as displayed in Fig. S3. The absorbance spectra revealed the 264 characteristic absorption peak of MB, which is located at 664 nm and the visual observation is 265

seen in Fig. 5(a). After 60 min of stirring in the dark, the intensity of the MB decreases slightly, 266 which confirms the adsorption of MB dye molecules on the surface of the photocatalyst. After 267 the UV light is switched ON, the intensity at 664 nm of MB decreases with time, indicating the 268 process of degradation of the dye molecule. The colour of the MB solution lightens with 269 increasing irradiation time and can be visually observed, as shown in Fig.  $5(a)$ . The C/C<sub>0</sub> plot 270 271 displayed in Fig. 5(b) shows the variation in MB degradation with different catalyst 272 concentrations. It is observed that with an increase in catalyst concentration, the kinetics of MB degradation increased. This is attributed to an increase in the number of active sites, which is 273 responsible for the generation of electron-hole pairs, generating more radicals, and improving 274 the degradation of MB dye. However, a maximum MB degradation is achieved for a catalyst 275 loading of 500 mg/L above which the MB degradation is slowed down. This is attributed to the 276 blocking of UV light into the MB dye solution by the catalyst, resulting in the saturation of 277 available photons and thus limiting the production of electron-hole pairs. 278



280 **Fig. 5.** (a) Visual appearance of MB dye solution at various time intervals in the presence of (CeGdHfPrZr)O<sub>2</sub> photocatalyst under UV irradiation. (b) A plot of  $C/C_0$  of MB degradation 281

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282 for various concentrations of the catalyst under UV irradiation and (c) its corresponding  $ln(C_0/C)$  plot (inset: rate constant). 283

Control experiments were also carried out to outline the importance of the catalyst and 284 UV-light source individually. Fig. S4 shows the absorption spectra of the MB dye solution in 285 the absence of a photocatalyst (Fig. S4(a)) and UV source (Fig. S4(b)). It is evident that in the 286 absence of any one, the MB degradation is not favorable, meaning both the UV light source 287 and photocatalyst are necessary for the photocatalytic reaction to carry on. 288

In general, the photocatalytic degradation of MB follows pseudo-first-order kinetics, 289 which is monitored by using a UV-visible spectrophotometer[53, 54]. Let  $[MB]_t$  be the 290 concentration of MB at time 't' and  $[MB]_0$  is the concentration of MB at time zero. Therefore, 291 on depends on [MB]. The<br>calculated rate constants<br>mg/L, 250 mg/L, 500 mg<br>g of 500 mg/L, a maximu the rate of the degradation reaction depends on [MB]. The rate constant k of the reaction is 292 given by  $k * t = ln \left( \frac{[MB]_0}{[MB]_t} \right)$ 293  $\frac{[M_{D10}]}{[MB]_t}$ . The calculated rate constants are 0.0042, 0.0054, 0.0078, and 0.0049 min<sup>-1</sup> respectively for 100 mg/L, 250 mg/L, 500 mg/L, and 1000 mg/L catalysts. In a 294 system with a photocatalyst loading of  $500$  mg/L, a maximum rate constant of 0.0078 min<sup>-1</sup> is 295 calculated, while a minimum value of  $0.0042$  min<sup>-1</sup> is obtained for a system with a 100 mg/L 296 catalyst loading. This being the case, we used a catalyst loading of 500 mg/L for further 297 investigations. 298

**3.1.2 Effect of methylene blue concentration** 299

300 Secondly, the effect of MB dye concentration was evaluated to study the impact of MB on photocatalytic reaction kinetics. The concentration of dyes in wastewater discharge varies 301 and are industry-dependent that are releasing the dyes. Therefore, it is necessary to study the 302 potential of photocatalysts on the impact of initial dye concentration. Moreover, the rate of 303 degradation is determined by the probability of hydroxyl radical formation on the catalyst 304 surface and the reaction of dye molecules with the hydroxyl radical. In this regard, several 305

initial dye concentrations such as 10 mg/L, 25 mg/L, 50 mg/L, and 100 mg/L were used, and 306 the photocatalytic degradation studies have been evaluated as shown in Fig. 6(a) and (b) and 307 Fig. S5. For lower MB concentrations, the degradation is faster, with a rate constant of 0.0093 308 min<sup>-1</sup>. This is probably due to the availability or generation of excess/sufficient hydroxyl 309 radicals by the photocatalyst, which increases the degradation kinetics of MB dye. 310



150 200 250<br>
ation on the photocatalytic<br>
iation and (b) its corresp **Fig.** 6. Effect of MB dye concentration on the photocatalytic reaction (a) Plot of  $C/C_0$  of MB 312 313 degradation under UV-light irradiation and (b) its corresponding  $ln(C_0/C)$  plot (inset: rate constant). (Catalyst: 500 mg/L) 314

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With an increase in MB concentration, the rate constant declines consistently, and a 315 value of 0.002 min<sup>-1</sup> is obtained for an MB concentration of 100 mg/L. Despite the fact that the 316 rate constant appears to decrease with increasing MB concentration, the quantity of MB dye 317 degraded in all cases is 0.0093 mg L<sup>-1</sup> min<sup>-1</sup>, 0.1975 mg L<sup>-1</sup> min<sup>-1</sup>, 0.19 mg L<sup>-1</sup> min<sup>-1</sup>, and 0.2 318 mg  $L^{-1}$  min<sup>-1</sup> respectively for 10 mg/L, 25 mg/L, 50 mg/L, and 100 mg/L of dye concentration. 319 The observed values seems intriguing since the amount of MB dye degraded is the same in all 320 cases except for the samples that contain10 mg/L of MB dye. Similar results have been 321 observed in other reported literature [55]. One probable explanation is the effect of 322 323 intermediary products on MB dye degradation. The intermediate products compete with MB dye for hydroxyl radicals[56]. A higher concentration of MB dye molecules around the active 324

sites of photocatalyst inhibited light penetration to the surface of the catalyst as a result of the 325 dye molecules. However, the explanation is not completely understood, so further research is 326 necessary to get a better understanding of the scientific rationale for the observed results. 327

**3.1.3 Effect of pH** 328

As discussed in the earlier section, similar to the MB concentration, the pH of industrial 329 330 water discharge is uncertain, and it is important to investigate the effect of solution pH on the photocatalytic degradation of MB dye. MB adsorption on photocatalysts is influenced by the 331 pH value since it depends on the surface state of the photocatalyst. The surface charge of the 332 catalyst is altered by pH values, which may either have a positive or negative impact on MB 333 degradation kinetics. Therefore, using 0.1 mol  $L^{-1}$  HCl or 0.1 mol  $L^{-1}$  NaOH, the pH of the 334 335 336 displayed in Fig. 7 (a), (b) and Fig. S6.





**Fig.** 7. Effect of pH on the photocatalytic degradation of MB dye solution (a) Plot of  $C/C_0$  of 338 MB degradation for various concentrations of the catalyst under UV irradiation and (b) its 339 corresponding  $ln(C_0/C)$  plot (inset: rate constant). (Catalyst: 500 mg/L; MB: 20 mg/L) 340

It is observed that the photocatalyst underperformed at acidic pH values, with the lowest 341 rate constant value of 0.0018 min<sup>-1</sup> (pH = 2.5). For a pH value of 4.4, a rate constant of 0.046 342 min<sup>-1</sup> is obtained. In contrast, at alkaline pH, increased photocatalytic degradation of MB dye 343

is observed. The faster reaction kinetics resulted in a rate constant of 0.011 min-1 for a pH of 344 10.5 while a rate constant of 0.0073 min-1 was obtained for a reaction pH of 8.6. The reason 345 for improved photodegradation of MB dye in alkaline conditions is due to the surface charges 346 present on the catalyst surface. In a basic pH environment, the photocatalyst surface is 347 negatively charged due to the presence of larger quantities of OH- ions. Similarly, under acidic 348 349 conditions, the nanoparticle is positively charged. As a result, the negatively charged 350 photocatalyst surface improves the adsorption of the cationic MB dye, increasing the photodegradation kinetics [57]. Another possible reason for improved photodegradation of MB 351 dye at a basic pH is owing to the generation of hydroxyl radicals in the alkaline environment. 352 There is a presence of an excess of OH<sup>-</sup> in the alkaline solution. On exposure to UV-light 353 irradiation, holes that are formed on the photocatalyst surface create hydroxyl radicals by the 354 is generated hydroxyl radiation<br>legradation of MB dye.<br>radation photooxidation of OH<sup>−</sup> [58]. This generated hydroxyl radical is expected to be the main 355 oxidizing species involved in the degradation of MB dye. 356

# **3.1.4 Mechanism of MB dye degradation** 357

To understand the mechanism of MB dye degradation by the HEO catalyst, it is 358 necessary to determine the band positions of the photocatalyst from the UV-visible absorption 359 spectrum. The band positions play a crucial role in determining the performance of the 360 photocatalyst [6]. With the help of Mulliken electronegativity, the valence band  $(E_{VB})$  and 361 conduction band ( $E_{CB}$ ) potentials can be calculated using the following equations (1) and (2) 362

363 
$$
E_{CB} = \chi - E^e - 0.5 E_g
$$
 (1)

$$
E_{VB} = E_{CB} + E_g \tag{2}
$$

where  $E^e$  is the energy of free electrons on the hydrogen scale (4.5 eV),  $E_g$  corresponds to the 365 bandgap of the material, which can be estimated from the Tauc plot (2.01 eV in this case) and 366

 $\chi$  denotes the electronegativity of oxides. Accordingly, equation (3) is used to calculate the 367 electronegativity of the overall semiconductor oxide<sup>[6]</sup>. 368

$$
369 \qquad \chi = \left(\ \chi A^a \ \chi B^b \ \chi C^c \ \chi D^d \ \chi E^e \chi F^f \right)^{1/(a+b+c+d+e+f)}
$$
 (3)

where a, b, c, d, e, and f are the number of atoms present in the compound, while A, B, C, D, 370 E, and F are the five metal cations and one anion present in the system. The electronegativity 371 of an individual atom can be found using (4) 372

373 
$$
\chi A = \frac{1}{2} (E_{EA}A + E_{IE}A)
$$
 (4)

where  $E_{EA}$  and  $E_{IE}$  are electron affinity energy and ionization energy, respectively. 374

Using the equations (1), (2), (3), and (4), the value of  $E_{VB}$  and  $E_{CB}$  is estimated to be +2.101 375 376 eV and +0.091 eV versus a normal hydrogen electrode (NHE), respectively.

I hydrogen electrode (NHI<br>
e mechanism for the de<br>
light over the photocata<br>
H<sub>2</sub>O<sub>2</sub>, electron-hole pairs a Fig. 8 shows the possible mechanism for the degradation of MB dye by HEO 377 photocatalyst. On exposure to UV light over the photocatalytic reaction solution containing 378 MB dye, HEO photocatalyst, and  $H_2O_2$ , electron-hole pairs are generated on the surface of the 379 photocatalyst, as displayed in equation (5). The electrons stay in the conduction band while the 380 holes stay in the valance band. Next, the holes in the valance band react with adsorbed water 381 and hydroxyl ions to produce 'OH radicals, as in equations (6) and (7). This reaction pathway 382 is possible because the calculated  $E_{VB}$  of the photocatalyst (+2.101 eV) is higher than that of 383 the redox potential of 'OH/OH'  $(+1.89 \text{ eV})$  [59]. 384

385 
$$
hv + HEO photocatalyst \longrightarrow e_{CB}^- + h_{VB}^+
$$
 (5)

$$
386 \t H_2O + h_{VB}^+ \longrightarrow {}^{\bullet}OH + H^+ \tag{6}
$$

 + ℎ ⎯⎯⎯⎯⎯⎯ • (7)

$$
388 \quad e_{CB}^- + H_2O_2 \longrightarrow + \, ^\bullet OH + OH^- \tag{8}
$$

Additional hydroxyl radicals are generated in the presence of  $H_2O_2$  by reacting with the 390 conduction band electrons, as indicated in equation (8). It should be noted that the generation 391 of superoxide radicals ( $^{\bullet}O_2^-$ ) is inhibited in the current system due to the lower E<sub>CB</sub> values of 392 the HEO catalyst. The redox potential of  $O_2/ O_2^-$  is -0.33 eV while the calculated E<sub>CB</sub> of the 393 394 photocatalyst is  $+0.091$  eV which is much lower than what is necessary for the reduction of  $O_2$ . 395 Therefore, the generated free radicals  $(°OH)$  interact with the MB dye, and the demethylation 396 process occurs by breaking the N-CH3 bond (equation (9)). A scavenging experiment was conducted with iso-propyl alcohol in order to confirm the involvement of hydroxyl radicals in 397 the degradation of MB dye [60]. Based on the scavenging results (Fig. 8(b)), it appears that 398 besonce of hydroxyl radica<br>ed in the photodegradation<br>nium ions, and sulfate ions MB is marginally degraded in the absence of hydroxyl radicals. Therefore, the hydroxyl radical 399 is the main reactive species involved in the photodegradation of MB dye. Finally, the MB dye 400 is degraded into  $H_2O$ ,  $CO_2$ , ammonium ions, and sulfate ions [61]. 401



402

**Fig. 8.** (a) Schematic displaying the mechanism of photocatalytic degradation of MB dye using 403 404 high-entropy (CeGdHfPrZr)O<sub>2</sub> oxide photocatalyst and (b) radical scavenging experiment on the photocatalytic degradation of MB dye solution . 405

To further investigate photocatalyst stability, recycling reactions were carried out for 406 the degradation of MB dye using the HEO catalyst. After the end of each cycle, the catalyst 407

was separated by centrifuge and washed with DI water, followed by drying at 80 °C and being 408 reused for the subsequent cycle. The degradation kinetics of MB decolorization during five 409 consecutive cycles under UV-light irradiation are displayed in Fig. 9(a) and its rate constant 410 plots are depicted in Fig. 9(b). 411



the photocatalytic degrad<br>JV-light irradiation and (the photocatalytic activity of **Fig. 9.** (a) Recyclability test for the photocatalytic degradation of MB dye solution using 413 (CeGdHfPrZr)O<sub>2</sub> catalyst under UV-light irradiation and (b) its corresponding  $ln(C_0/C)$  plot 414 (inset: rate constant). (Catalyst:  $500 \text{ mg/L} \cdot \text{MB} \cdot 20 \text{ mg/L}$ ) 415

412

416 According to the results, the photocatalytic activity of the HEO catalyst was stable over 417 the course of five cycles. The rate constant of MB decolorization decreased slightly from 0.0083 to 0.0070 min-1 , indicating that the HEO catalyst had good stability. To probe the 418 structural properties of the HEO catalyst, XRD was performed on the recycled photocatalyst 419 after 5 cycles, and the X-ray pattern is displayed in Fig. S7. The XRD results confirm the 420 stability of the catalyst without displaying any structural transformation during the 421 photocatalytic studies. In addition, FTIR spectroscopy has been carried out to investigate the 422 surface functional groups present on the surface of the recycled photocatalyst, and the plot is 423 424 displayed in Fig. S8. By comparing the FTIR spectra of the photocatalyst before and after MB degradation, it is clear that the catalyst has retained its structure and there has been no 425 significant change in the functional groups present in the catalyst. As a result, it is evident that 426

the synthesized high-entropy oxide catalyst exhibits high stability and is capable of being 427 recycled numerous times without losing any photocatalytic activity. 428

#### 429 **4. Conclusion**

A simple hydrothermal synthesis route was used to synthesize high-entropy (CeGdHfPrZr) $O<sub>2</sub>$ 430 oxide nanoparticles. The synthesized HEO exhibits a single-phase cubic fluorite crystal 431 structure with a mean crystallite size of 5.4 nm. A higher lattice strain is found in the HEO 432 system due to the presence of different-sized metal cations. The MB degradation kinetics were 433 evaluated by varying the concentration of the photocatalyst, MB dye, and pH. MB degradation 434 studies indicate better photocatalytic activity by the synthesized HEO photocatalysts. 435 Moreover, the degradation kinetics are dependent on several experimental conditions, such as 436 tration, and solution pH. I<br>yst concentrations, which<br>yst performed better at a<br>proving MB dye degrad 437 catalyst concentration, dye concentration, and solution pH. Poor MB dye degradation kinetics are observed at higher photocatalyst concentrations, which limits photon penetration into the 438 reaction solution. The photocatalyst performed better at a basic pH as a result of excess 439 hydroxyl radical generation, improving MB dye degradation. Moreover, based on the 440 calculated EVB and ECB values, hydroxyl radicals are responsible for the degradation of MB 441 dye, which is confirmed by the hydroxyl radical scavenging study. In addition, the HEO 442 443 catalyst material was found to be stable and reused for several degradation cycles without decreasing the efficiency of the photocatalyst. (CeGdHfPrZr) $O_2$  was found to be a promising 444 photocatalyst for the degradation of organic pollutants under UV irradiation. The use of novel 445 high-entropy oxides in water treatment applications is an attractive one. 446

447

# **CRediT authorship contribution statement**

**Mariappan Anandkumar:** Conceptualization, Data curation, Formal analysis, Investigation, 448 Methodology, Validation, Writing – original draft, Writing – review & editing, Visualization. 449 **Kannan P.K.:** Data curation, Formal analysis, Investigation, Validation, Writing – review & 450

- editing. **Shanmugavel Sudarsan:** Investigation, Writing Review & Editing. **Evgeny**  451
- **Trofimov:** Conceptualization, Funding acquisition, Project administration, Supervision 452

## **Declaration of Competing Interest** 453

There are no conflicts to declare. 454

#### **Acknowledgements** 455

The study was supported by the Russian Science Foundation Grant No. 22-23-00243, 456 https://rscf.ru/project/22-23-00243/ 457

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